

Research Article

Enhancing the Cu (II) Removal in Water Using Functional Hybrid Membranes

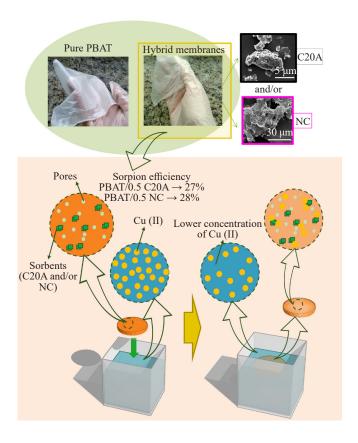
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Abstract: This work developed hybrid EIPS/NIPS membranes of poly (butylene adipate-co-terephthalate) (PBAT) with nanocurcumin (NC) and/or Cloisite 20A (C20A). The dispersed phases were characterized by dynamic light scattering (DLS), ζ potential, X-ray diffractometry (XRD), scanning electron microscopy (SEM), and Fourier-transform infrared spectroscopy (FTIR), while XRD, SEM, FTIR, mechanical properties, contact angle, and copper sorption were evaluated for the composite membranes. DLS analysis indicated that the dispersed phases present a nanometric size distribution; ζ potential measurements showed low electrostatic stability, explaining the agglomeration effects observed. Pure PBAT membranes presented macro-pores throughout their structure, which showed a slight size reduction with the inclusion of NC and C20A. The membranes' mechanical performance was affected by the presence of the pores that functioned as stress-concentrating defects, and the inclusion of the dispersed phases increased the elastic modulus and tensile strength, especially for PBAT/0.5% C20A, with values of 18.7 and 8.9% higher than those of pure PBAT, respectively. Hybrid EIPS/NIPS membranes showed a hydrophilic nature, with all the contact angle measurements lower than 90°. The sorption tests using a high-concentration copper solution (1,000 mg/L) showed a removal of around 25%. These results highlighted the development of new eco-friendly membranes using NC and C20A as dispersed phases with the potential to remove toxic elements from water.

Graphical abstract



Keywords: membranes, water treatment, curcumin, cloisite 20A

1. Introduction

The presence of heavy metals in the environment is a global problem that poses severe threats to marine life and human health due to their toxicity, non-biodegradability, and bioaccumulation. Heavy metals refer to elements with atomic densities above 4.5 g cm⁻³ and are derived from anthropogenic sources, such as petroleum, metal smelting, and electroplating. A more up-to-date nomenclature can refer to heavy metals and other harmful elements as potentially toxic elements (PTEs) since, above a specific concentration, they can be harmful to the ecosystem and living organisms. There are several PTEs related to anthropogenic activity, such as antimony (Sb), arsenic (As), beryllium (Be), cadmium (Cd), fluoride (F), gold (Au), lead (Pb), mercury (Hg), molybdenum (Mo), nickel (Ni), rubidium (Rb), selenium (Se), silver (Ag), uranium (U), vanadium (V), zinc (Zn), and copper (Cu).

As a transition metal, copper can be found in nature in different oxidation states, such as a monovalent and bivalent ion, Cu (I) and Cu (II), respectively. While Cu (I) is absorbed into the intestine through food intake, Cu (II) can originate from water sources and may not be absorbed, increasing the concentration of copper in the human organism. Due to its industrial applications, Cu (II) can also be found in high wastewater concentrations. However, the permissible limit for copper ions in industrial effluents is reported as 1.3 and 2.0 ppm by the World Health Organization and USEPA, respectively. Above these limits, the literature reports toxic effects on human health, such as carriers of the gene for Wilson disease, gastrointestinal effects, metabolic disorders, and even death.

Over the last decades, different techniques have been used to remove PTEs, such as chemical precipitation, ¹² ion exchange, ¹³ ion flotation, ¹⁴ coagulation/flocculation, osmosis, and electrochemical removal. However, these methods have disadvantages, such as erratic removal of metal ions, creation of toxic sludge, or high energy consumption, ¹⁵

and, thus stimulate the search for cheaper and high-efficiency modern technologies. Among the PTE removal trends, adsorption is highlighted as a promising and efficient technique. ¹⁶ Few authors have reported combining the adsorption process in membranes, aiming for higher efficiency and metals removal selectivity, where sorbents can be used as fine particles to increase the interface area available for adsorption. ¹⁷⁻¹⁹ Several materials can be used to prepare the membranes, such as poly (vinylidene fluoride)/polyaniline, polyethylene, polyacrylonitrile, polyvinyl alcohol, and poly (N, N-dimethylacrylamide-co-2-hydroxyethyl methacrylate), ²⁰ and also 2D materials such as boron nitride, ²¹ graphene and MXene. ²³ However, due to environmental concerns, biodegradable polymers, e.g., cellulose, ²⁴ polylactic acid, ²⁵ and polybutylene adipate-co-terephthalate, ²⁶ have been widely used for membranes.

Inorganic adsorbents can remove PTEs from water, and the literature covers examples of inorganic-based adsorbents, such as layered double hydroxide, ²⁷ metal-organic frameworks, ²⁸ silica-based titanate, ²⁹ and clays. ³⁰⁻³² Montmorillonite (MMT) is a clay and one of the most studied inorganic materials for this application due to its high cation exchange capacity, natural electronegativity, controllable interlamellar spacing, and green and non-toxic properties. ³³ The interlamellar space can be functionalized to enhance MMT potential application further and offer newly available interactions. The insertion of hydrogenated ammonium organic modifiers deserves a highlight since MMT modified with those modifiers is known as Cloisite. ³⁴ Mohajeri and co-workers studied nano Cloisite Na⁺ to remove Pb (II) and Cu (II), obtaining 40% and 16% adsorption efficiency, respectively. ³¹ Another type of Cloisite, the Cloisite® 20A, presents a potential application for heavy metal removal in wastewater, which has not been extensively explored in literature for Cu (II) removal. Although it is not commonly used to remove PTEs, justifications for using Cloisite® 20A can be found in the literature. ³⁵

Besides inorganic-based materials, organic-based materials such as cellulose, ³⁶ natural fibers residues, ³⁷ chitosan, ³⁸ and curcumin³⁹ also can remove PTEs from aqueous media. Curcumin is a polyphenolic organic compound from turmeric (Curcuma longa) of formula C₂₁H₂₀O₆ and is a biomaterial little studied for this purpose but with great application capacity due to its symmetric molecular structure and terminations of hydroxyl and ketone functional groups. ⁴⁰⁻⁴¹ Ferrari, Xi, and co-workers related that curcumin could have a strong chelation effect on metal ions. ⁴²⁻⁴³ Still, the complexes formed between curcumin-metal ions lack further investigation of their complexes' properties. Both clays and curcumins can be dispersed phases in polymeric matrix membranes to remove PTEs from contaminated water. Moradi and co-workers used boehmite clay particles modified with curcumin in poly (ether-sulfone) membranes obtained by phase inversion for retention of multiple bivalent ions such as Cu (II), where the addition of 0.5% loading of clay and curcumin combined decreased the content of the contaminants in water. ⁴⁴

According to Hashim and co-workers, hybrid materials are nanocomposites with at least one element at a nanoscale. ⁴⁵ Therefore, this article is a follow-up to a previous study ³⁵ and investigated the adsorption-hybrid membranes prepared using PBAT as a polymeric matrix, nanocurcumin (0.5% m/m), and nanoclay Cloisite® 20A (C20A (0.5% m/m) as sorbents, aiming their potential application in reducing PTEs concentration in wastewater. Besides, the efficacy of the hybrid EIPS/NIPS membranes was investigated in reducing the Cu (II) concentration in a solution simulating an industrial effluent.

2. Materials and methods

2.1 Materials

PBAT, commercially named Ecoflex® 87 F Blend C1200 density of 1.25 g cm⁻³ at 23 °C, (BASF SE, Germany), was supplied by BASF SE (São Paulo, Brazil). Chloroform (P.A.-ACS), sodium chloride (NaCl) (P.A.-ACS), ethyl alcohol (99, 5% of purity), copper nitrate II trihydrate (Cu(NO₃)₂)·3H₂O) (P.A.-ACS), and nitric acid (HNO₃) (70% of purity) were purchased from Synth (São Paulo, Brazil). The modified montmorillonite nanoclay BYK Cloisite® 20A Nanoclay (C20A) was purchased from Songhan Plastic Technology Co., Ltd. The turmeric powder (or fresh turmeric) derived from curcumin was purchased from a local supermarket (Jundiaí-SP).

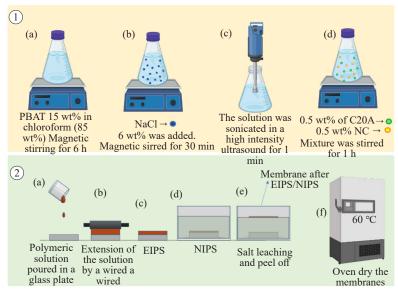
2.2 Methods

Curcumin preparation:

Firstly, 40 g of turmeric powder was added to 200 mL of ethanol and kept under magnetic stirring for 1 h to obtain curcumin. Then, the mixture was vacuum-filtered and dried at 30 °C for 24 h. Nano-sized curcumin (NC) structures were obtained by mechanical ball milling using a rotational mill (Model Q298, brand QUIMIS, 110 V, 60 Hz, 25 W, 150 rpm rotation on the drive roller). The grinding method was adapted from Jha and co-workers, tusing a metal pot (300 mL) attached to a 1 L porcelain jar model 21 A (Chiarotti). 200 g of alumina beads (10 mm diameter) and 10 g of curcumin (ratio 20:1 m/m) were added to the metal pot. Before grinding, the curcumin was dried at 30 °C for 24 h. The milling process took place for 24 h at room temperature.

Polymer solution preparation:

A polymeric solution of PBAT 15 wt% in chloroform (85 wt%) was prepared under constant magnetic stirring for 6 h. A saturated aqueous sodium chloride (NaCl) solution (porogen with a volume corresponding to 6 wt% relative to the PBAT 47) was added. The mixtures were magnetically stirred for 30 min and then subjected to high-intensity ultrasound for 1 min (Sonics Vibra-cell VCX 500-VCX 750 sonicators) under the operating conditions: energy of 92,000 J and amplitude of 50%. The solutions with phase insertion of C20A and/or NC had an extra step in the mixing process before ultrasonication by adding the pre-set proportions of the dispersed phase(s), e.g., 0.5 wt% of C20A and/or 0.5 wt% NC. After adding the dispersed phase(s), the mixture was stirred for 1 h.



All the membranes are named EIPS/NIPS membranes, and the membranes with NC and/ or C20A are named hybrid EIPS/NIPS membranes

Figure 1. Illustrative scheme of methodology preparation for the functional hybrid EIPS/NIPS membrane, (1a-c) preparation of PBAT solution and (1d) PBAT membrane with C20A and/or NC loading; (2a-e) EIPS/NIPS membranes structure obtaining and (2f) the final drying step

Membrane preparation by EIPS/NIPS:

The polymer solution previously described (with porogen and with or without dispersed phase(s)) was applied to glass plates (13×18 cm) and spread out using a 200 µm Erichsen wire (200 µm). The procedure was done with 1 min drying time for partial solvent evaporation (EIPS). After evaporation, the plates were immersed in polyethylene trays in a non-solvent (NIPS) (distilled water) coagulation bath to induce phase separation. Then, the membranes were separated from the glass plates and remained in a distilled water bath for 24 h for salt leaching, with a subsequent dry process in an oven at 60 °C for 6 h.⁴⁷ The EIPS/NIPS membranes were obtained, as shown in Figure 1, and the compositions were based on a previous study³⁵ and the literature.⁴⁴ The composition of the polymer solution and dispersed phases (NC and C20A) are described in Table 1.

Table 1. Composition of EIPS/NIPS membranes developed, with the percentages of PBAT solution, C20A, and/or NC powders

Membrane	PBAT solution (%)	Cloisite® 20A (C20A) (%)	Curcumin (NC) (%)
PBAT	100.0	-	-
PBAT/0.5 C20A	99.5	0.5	-
PBAT/0.5 NC	99.5	-	0.5
PBAT/0.5 C20A + 0.5 NC	99.0	0.5	0.5

2.3 Characterization

2.3.1 Dynamic light scattering (DLS) and ζ potential

The Dynamic light scattering (DLS) recorded the average particle diameter of the adsorbents (C20A and NC). It was done with a stable 90° scattering angle using Zetasizer Nano-ZS (Malvern Panalytical Ltd., Malvern, U.K.). The adsorbents were prepared with 1 mL aliquots and diluted in 100 mL distilled water. The tests were done in triplicate. The ζ potential used the same equipment and sample preparation.

2.3.2 X-ray diffractometry (XRD)

The adsorbents (C20A and NC) and EIPS/NIPS membranes were analyzed by X-ray diffraction measurements using a D8 Focus diffractometer (Bruker AXS Advanced X-ray Solutions GmbH, Karlsruhe, Germany), operating at 40 kV and 40 mA. It used monochromatic CuK α radiation ($\lambda = 1.54056$ Å) at a step width of 0.01° and counting time of 100 s at each 0.5°, from $2\theta = 10^{\circ}$ to 60° .

2.3.3 Scanning electron microscopy (SEM)

The adsorbents (C20A and NC) and EIPS/NIPS membrane samples (surface and cross-section) were coated with a 25 nm thick gold layer using Sputtering Leica EM ACE 200 (Leica Microsystems, Wetzlar, Germany). The cross-section membrane samples were obtained by fracturing the membrane after nitrogen immersion. Micrographs were performed using a microscope FEI Quanta 250 (Thermo Fisher Scientific, Hillsboro, Oregon, USA), using an accelerating voltage of 10 kV and a spot size of 4 nm.

2.3.4 Fourier-transform infrared spectroscopy

The adsorbents (C20A and NC) and EIPS/NIPS membranes were analyzed by Fourier-transform infrared spectroscopy (FTIR) using a Frontier 94.942 (PerkinElmer Inc, Massachusetts, USA) with attenuated total reflectance (ATR) mode in a range of 500-4,000 cm⁻¹, 32 scans, and spectral resolution of 4 cm⁻¹.

2.3.5 Mechanical testing

Tensile tests analyzed the EIPS/NIPS membranes using an Instron 3367 (Instron®, USA Norwood, USA). All experiments were conducted according to the ASTM D638-14 standard. Five specimens were evaluated with a load cell of 50 N and 50 mm min⁻¹ test speed.

2.3.6 Contact angle (CA)

The EIPS/NIPS membranes were analyzed by contact angle using a dynamic contact angle tensiometer (SEO) model Phoenix 300 (Kromtech Alliance Corp., London, UK), with three phases (air/water/sample). The static contact angles (θ_{est}) were determined according to the Young-Laplace equation by observing a water droplet of 4.00 μ L deposited on the membrane surface. The contact angle values were collected in triplicates for 5 minutes, in 3-second intervals of each measurement, and a mean value of the collected values was considered.

2.3.7 Copper adsorption

A Cu (II) solution containing 1,000 ppm of Cu (II) was prepared to simulate the effluent from a plating industry for the adsorption experiments. 48 Cu(NO₃)₂·3H₂O was dissolved in a solution of 0.01 M HNO₃, and a sodium hydroxide solution of 0.01 M NaOH was used for pH 5 correction. The pH was set to 5 due to precipitation and low stability of aqueous solutions with pH > 5. 49 Once the solution was prepared, 0.1 g of each EIPS/NIPS membrane was placed in a plastic tube and mixed with 50 mL of Cu (II) solution, on the other hand, an adsorbent dosage of 2 g/L. 50 Adsorption measurements in triplicate were performed from the beginning (t = 0) and at times (t) of 10 min, 45 min, and 1,440 min.

Evaluation of Cu (II) concentration for adsorption occurred using Cary 50 UV-vis spectrophotometer (Varian) at the wavelength of 305 nm for Cu (II). ⁵¹⁻⁵² The adsorption efficiency was calculated using Equation 1, where C_o (mg/L) is the initial concentration of the metal ion in the wastewater before adsorption, and C_f (mg/L) is the final concentration of the metal ion in the wastewater after adsorption.

Efficiency% =
$$\frac{C_o - C_f}{C_o} \times 100\%$$
 (1)

3. Results and discussion

3.1 Dynamic light scattering (DLS) and ζ potential

Dynamic light scattering (DLS) was used to evaluate the distribution profile and hydrodynamic radius (RH) of C20A and NC, considering these nanomaterials to be spheres. The C20A showed a trimodal distribution profile (Figure 2), with average peak values around 10.1, 342, and 5,560 nm, indicating a polydisperse and heterogeneous sample. The first two peak values are characteristic of nanoclays, as reported in the literature, while the higher radius values (5,560 nm) indicate lamellae agglomerations due to van der Waals bonds. According to Zdiri and co-workers, although individual nanoclays have nanometric dimensions, MMT aggregates have micrometric dimensions, between 0.1-10 μm, consisting of lamellae in random directions.⁵³ The presence of agglomerates justifies using low C20A contents in the membranes, aiming to achieve better dispersion and adequate performance for application in water treatment.

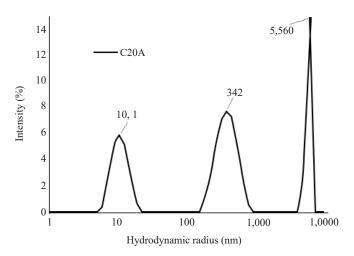


Figure 2. Distribution profile and hydrodynamic radius (RH) of C20A

The nanocurcumin DLS results showed a broad peak, and the wide distribution is associated with the mechanical grinding process that results in irregularly shaped and amorphous particles.⁵⁴ It was observed that two peaks were located at 255 and 4,150 nm. The ball milling process is a top-down mechanical method of green synthesis that allows the conversion of microparticles into nanometric dimensions,⁵⁵ as observed in Figure 3.

The ζ potential was used to evaluate the electrostatic stability of the nanoparticles, and the samples showed ζ potential values of -11.5 and -3.2 mV for C20A and NC, respectively. According to the literature, particles with high ζ potential values in the module present higher electrical stability, while particles with lower ζ potential could present agglomeration tendency. Since the dispersed phases have low ζ potential, they can present agglomeration effects, corroborated by DLS results, with C20A lamellae clusters and poor NC dispersion.

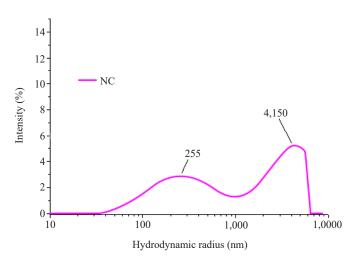


Figure 3. Distribution profile and hydrodynamic radius (RH) of NC

3.2 X-ray diffractometry

The XRD technique was used to determine the materials' crystalline pattern, and Figure 4 shows the obtained diffractograms.

C20A showed two broad peaks between $2\theta = 0\text{-}10^\circ$, at i) 3.6° and ii) 7.1°, where peak i) refers to the crystallographic plane (001), and peak ii) refers to the crystallographic plane (002).⁵⁷ The peak i) showed an interplanar distance (d) of 24.2 Å and is assigned to the montmorillonite-modified layers,⁵⁸⁻⁶¹ corroborating the organomodified nature of Cloisite® 20A. The peak ii) showed d = 12.4 Å, associated with the layers that did not suffer the influence of the organic modifier.⁶² Other peaks resulting from the modification of MMT were also seen at iii) 19°, iv) 24° and v) 35° (Figure 4a). According to Patiño-Soto and co-workers, the peak at 3.6° (Figure 4b) is the most characteristic for organomodified clays, corresponding to the original basal reflection peak for Cloisite® 20A.⁶⁴ Abdolrsaouli and co-workers reported the second peak of Cloisite® 20A at 8.2°, reflecting the smaller interplanar distance, i.e., more excellent packing of the MMT layers.⁵⁹ The crystallinity index (CI) of the C20A was 85.6%, which agrees with literature reports.⁶⁵

Figure 4c and d-e demonstrate the curcumin XRD patterns before and after the milling process, respectively. Figure 4c shows a broad band with two peaks at 17.4 and 22.1°, highlighting the amorphous characteristics of curcumin associated with the turmeric organic nature. After the mechanical milling (Figure 4d-e), the amorphous character of the sample was accentuated due to the ball grinding friction. In the nanometric particles, there is an increase in surface defects and random bonds to facilitate adsorption, as illustrated in Figure 4f. Zhang and co-workers reported an amorphization of curcumin nanoparticles milled for 24 h at 157 rpm. According to Qiu and co-workers, the mechanical energy in the grinding process can disorganize crystalline areas and result in amorphization, increasing samples adsorption capacity due to increased particle surface area. However, the agglomeration process can mask the decrease in particle size after long grinding times. The authors reported that the relationship between amorphization and particle agglomeration still requires further studies to understand the linkage of factors. Other authors reported increased particle sorption capacity after amorphization.

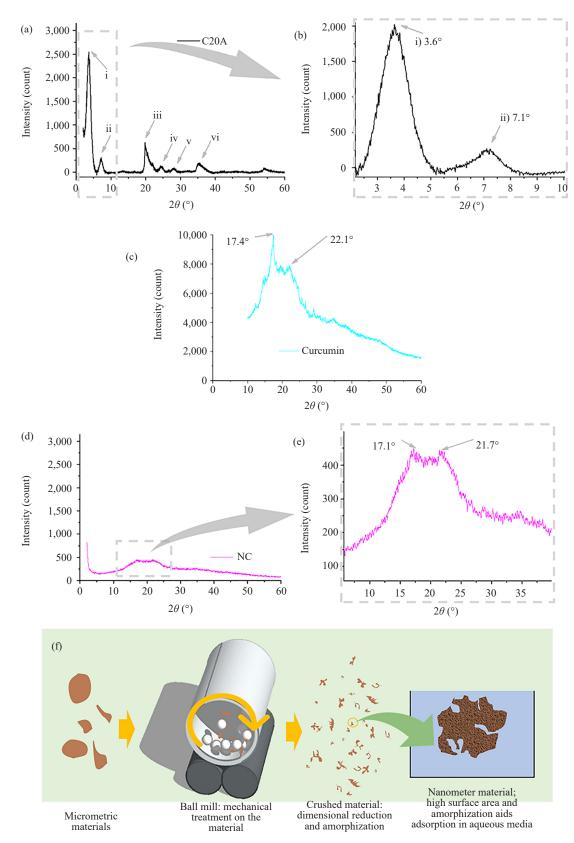


Figure 4. Diffractometer patterns relating to C20A at (a) $2\theta = 0.60^{\circ}$, and (b) zoom showing the $2\theta = 0.10^{\circ}$ region, (c) unground curcumin, and to NC at (d) $2\theta = 0.60^{\circ}$ and zoom at (e) $2\theta = 3.38^{\circ}$, (f) illustrates the mechanical grinding of curcumin

The lack of atomic periodicity of adsorbent materials also influences diffusion processes involving adsorption. The NC diffractometric profile is like the nanometric curcumin well presented in the literature. The NC diffractometric profile is like the nanometric curcumin well presented in the literature. Should be a broad band centered at $2\theta = 17^{\circ}$, which was used to calculate the CI value of 30.6% in this work.

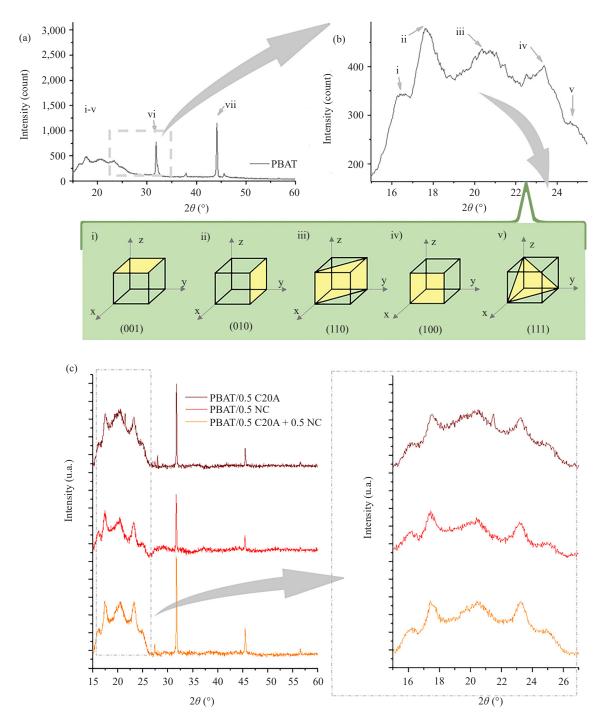


Figure 5. Diffractometer pattern for (a) pure PBAT membrane, (b) zoom for better visualization between 10-26°, and graphical representation of the Miller Indices. (c) hybrid EIPS/NIPS membranes ($15^{\circ} < 2\theta < 60^{\circ}$) and zoom for best visualization 15-27°

The XRD was also used to investigate the developed EIPS/NIPS membranes, as shown in Figure 5.

Figure 5a shows a standard PBAT diffractogram with characteristic peaks between $2\theta = 15\text{-}25^\circ$ (Figure 5b, reflections from i to v). ⁷⁸⁻⁸¹ Kashi and co-workers described that the peaks in the mentioned range correspond to the characteristic peaks of the PBAT due to the basal reflections in the Miller index planes $2\theta = 16.3^\circ$ (001), $2\theta = 17.6^\circ$ (010), $2\theta = 20.3^\circ$ (110), $2\theta = 23.3^\circ$ (100), and $2\theta = 24.6^\circ$ (111). Figure 5c shows the XRD patterns of the membranes, which showed high similarity with the pure polymer.

The diffractograms with lower peak intensity in the hybrid EIPS/NIPS membranes show more excellent exfoliation of C20A and dispersion of NC along the membrane weight, despite the tendency to agglomerate seen in DLS and ζ potential. The nanometer size of C20A may facilitate the dispersion and interface reinforcement/matrix, influencing the crystallinity and enabling a more outstanding adhesion of the nanoparticles of PBAT films for better mechanical properties. 82-83

3.3 Scanning electron microscopy

The micrographs obtained by SEM of the surface and cross-section of EIPS/NIPS membranes are shown in Figure 6.

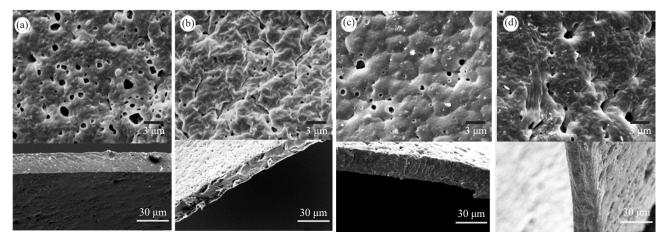


Figure 6. SEM micrographs of EIPS/NIPS membranes surface: (a) PBAT, (b) PBAT/0.5 C20A, (c) PBAT/0.5 NC, and (d) PBAT/0.5 C20A + 0.5 NC

The results demonstrated that the surface of the pure PBAT membrane presents pores with a more homogeneous cross-section (less surface roughness) than the hybrid membranes. For PBAT/0.5 C20A, it was noted that adding nanoclay decreases the pore diameter by changing the surface, leaving it more irregular and its cross-section, while for PBAT/0.5 NC, the surface becomes less rough. Such differences between the dispersions of the phases may result from different interactions with the matrix since C20A tends to exfoliate when in polymer solutions. The PBAT/0.5 C20A + 0.5 NC sample has more evidence of a dispersed phase due to its irregularity, agglomerations, and higher additive fractions.

Table 2 shows the average diameter and maximum pore size measurements of the EIPS/NIPS membranes and the average dry thickness measurements (cross-section after rupture).

Table 2. Average (μm)) and maximum pore size	e (μm); and EIPS/NIPS mei	mbranes cross-sectional thickness	s (μm) (dry thickness)

Membrane	Average pore size (μm)	Max. pore size (μm)	Transversal section (μm)
PBAT	0.6 ± 0.5	4.6	22.9 ± 0.6
PBAT/0.5 C20A	0.5 ± 0.2	1.7	15.4 ± 1.2
PBAT/0.5 NC	0.5 ± 0.4	2.7	23.6 ± 1.5
PBAT/0.5 C20A + 0.5 NC	0.4 ± 0.3	2.0	26.5 ± 0.7

Compared to pure PBAT, all hybrid EIPS/NIPS membranes showed smaller average pore diameters ranging from 0.4 to 0.5 µm. Barbosa and co-workers obtained pores ranging from 0.25 to 0.6 µm in PBAT membranes with 6% NaCl as porogen. The pores of the membranes for both preparation methods are classified as macropores because they exceed the size of 50 nm (according to the standard of the International Union of Pure and Applied Chemistry (IUPAC)). Although PBAT/0.5% C20A membrane has lower dry thickness than pure PBAT (15.4 and 22.9 µm, respectively), most EIPS/NIPS membranes have thicknesses above 21 µm. According to Liu and co-workers, Young's modulus and tensile strength limit depend on a membrane's thickness. Lima and co-workers observed that increasing the thickness of elastomeric membranes decreased both the wettability seen by the contact angle and the hardness of the membranes.

3.4 Fourier-transform infrared spectroscopy

The FTIR analysis was also used to investigate the developed EIPS/NIPS membranes, and the obtained spectra are presented in Figure 7.

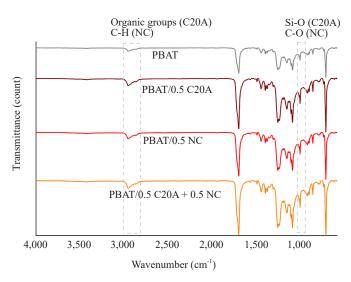


Figure 7. FTIR spectra EIPS/NIPS membranes, highlighting the prominent peaks associated with C20A and NC

The prominent peaks associated with PBAT structure are observed at 3,065 cm⁻¹ (C-H stretching), 2,958, 2,875 cm⁻¹ (CH₃ and CH₂ stretching), 1,710 cm⁻¹ (C = O bond of ester group), 1,410 cm⁻¹ (bending mode of O-CH₂), 1,366 cm⁻¹ (bending mode of CH₂), 1,267 cm⁻¹ (C-O group) and 726 cm⁻¹ (deformations of the benzene ring). All EIPS/NIPS membranes presented these characteristics peaks.⁸⁸⁻⁸⁹

Regarding the hybrid EIPS/NIPS membranes, the bands found for NC are commonly observed at 3,312 cm⁻¹ (OH stretching of the phenolic group), 2,920-2,851 cm⁻¹ (C-H stretching), 1,638 cm⁻¹ (C = C and C = O stretching), 1,241 cm⁻¹ (angular deformation in the C-O-C plane), 1,101 cm⁻¹ (C-O stretching), and 1,018 cm⁻¹ (C-O-C stretching). The bands corresponded to the chemical composition of phenolic groups for NC,⁹³ and these bands are essential for the sorption process of contaminant metal cations, as they can complex bivalent cations due to the direct binding of curcumin molecule with the ion or by hydrogen bond formation.⁹⁴ For C20A, the prominent peaks are observed at 3,632 cm⁻¹ (O-H stretching), 2,921, 2,849 and 1,475 cm⁻¹ (C-H stretching of alkylammonium groups), 1,639 cm⁻¹ (O-H angular deformation), 1,004 cm⁻¹ (Si-O vibration), 917 cm⁻¹ (Al-OH-Al deformation) and 725 cm⁻¹ (Si-O stretching). Similarly, the silicates, aluminate, and alkylammonium groups are relevant to promote the complexation and removal of contaminants. The peaks of C20A and NC can be seen in Figure 8.

Comparing the FTIR spectra of hybrid EIPS/NIPS membranes, it was observed that they were mainly related to PBAT typical peaks, which could be attributed to i) the overlapping of some peaks of PBAT that are observed in similar regions of those from NC and C20A; ii) the low concentration of NC and C20A used in the membrane's formulation.

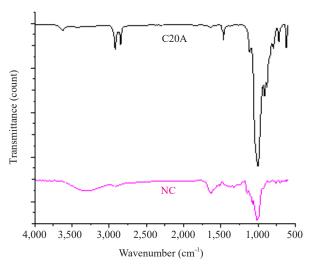


Figure 8. FTIR spectra of C20A and NC

3.5 Mechanical testing

The mechanical test was employed to investigate the mechanical properties of EIPS/NIPS membranes (Figure 9).

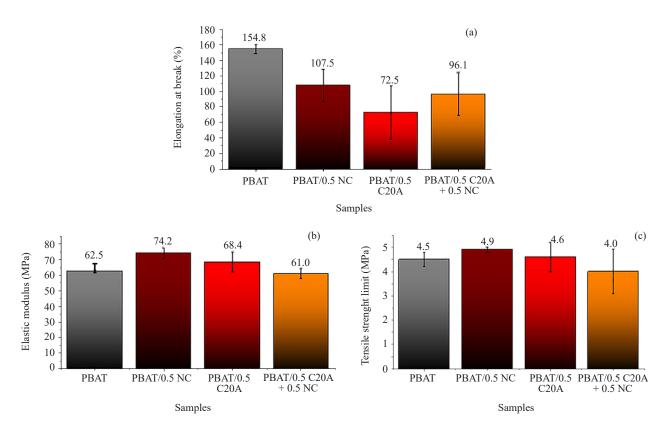


Figure 9. Mechanical properties of pure PBAT and hybrid EIPS/NIPS membranes, being (a) elongation at break (%), (b) elastic modulus (MPa), and (c) tensile strength limit (MPa)

The strain of a membrane is the elongation at break (%), which measures the flexibility, i.e., the ability of the film

to deform before cracking and eventual rupture. ⁹⁷ In Figure 9a, it was observed that the pure PBAT membrane obtained the highest elongation at break (154%). In the literature, the elongation at break for PBAT is higher, with records of up to 508%. ⁹⁸ The hybrid EIPS/NIPS membranes are smaller, which is associated with the fact that it is a porous membrane that may have affected the deformation since the pores may have acted as stress-concentrating defects that facilitate rupture. ⁴⁷ The dispersed phase can reinforce the PBAT structure, increasing its rigidity and reducing the membrane's elongation. Both C20A and NC influenced this property, and among the hybrid EIPS/NIPS membranes, the one with the highest strain was PBAT/0.5 C20A (107.5%), and the lowest value was for the PBAT/0.5% NC membrane (72.5%). Thus, PBAT/0.5 C20A + 0.5 NC had an intermediate value compared to membranes with only one dispersed phase (96.1%).

A polymer's stiffness is determined by the elastic modulus, obtained by the slope of the linear portion of the stress-strain distribution for a polymer under compression. It is a widely popular metric for choosing materials with superior mechanical attributes. Figure 9b shows that membranes with only one dispersed phase added had a higher elastic modulus. Correia and co-workers also observed this behavior in PBAT membranes with charcoal addition, showing that adding dispersed phases increased membrane integrity. In this work, compared to pure PBAT elastic modulus of 62.5 MPa, PBAT/0.5 C20A had an increase of 18.72% (74.2 MPa), and PBAT/0.5 NC had an increase of 9.44% (68.4 MPa). This fact corroborates the increase in IC seen in the XRD analysis, which increases the membrane's stiffness.

In Figure 9c, it was observed that the insertion of C20A in the PBAT/0.5 C20A (with 4.9 MPa) was the one that most favored the increase in tensile strength (8.9%) when compared to that of pure PBAT (4.5 MPa), followed by the PBAT/0.5 NC (4.6 MPa). Crystalline particle loading in polymeric matrix increases the stiffness of materials and thus can influence their tensile strength. The PBAT/0.5 C20A + 0.5 NC membrane, on the other hand, showed no increase and had the lowest tensile strength limit (4.0 MPa), which may be an indication that both dispersed phases (C20A and NC) in the matrix did not favor the matrix/reinforcement interaction. Ludwiczak and co-workers studied the mechanical behavior of MMT in a PBAT and PLA polymeric matrix. The authors concluded that the role of clay as an active interfacial modifier in a polymeric composite still needs further research to understand the effect of the dispersed phase on material stiffness and deformation. Thus, the insertion of C20A and NC both separately and simultaneously is still a research object that is not fully understood. Research object that is not fully understood.

3.6 Contact angle

The contact angle (CA) is an essential measurement to investigate the surface behavior of the EIPS/NIPS membranes and their interaction with water. Figure 10 shows the dynamic CA over 5 min (Figure 10a), the representation of different CA classifications (Figure 10b), and the average CA (Figure 10c).

The dynamic measurement of the CA can investigate the interaction of the water droplet with the membrane surface as a function of time. Figure 10a showed that EIPS/NIPS membranes decreased in CA over the test time (5 min). It was also observed that CA values were lower than 90° throughout the assay. As illustrated in Figure 10b, hydrophilic materials have CA surfaces less than 90°, and hydrophobic materials have CA above 90°. Therefore, the EIPS/NIPS membranes obtained are hydrophilic. That behavior was expected due to the hydrophilic nature of the biopolymer PBAT. Yasin and co-workers observed that the water droplet on poly (vinyl alcohol) membranes with chitosan particles exhibited a decrease in CA concerning time, i.e., the biopolymer matrix surface with a dispersed phase exhibits a hydrophilic nature. Lin and co-workers also observed decreased CA with time for a polyhydroxybutyrate (PHB)/PBAT membrane. Furthermore, the porous surface of the EIPS/NIPS membranes seen in the SEM micrographs may have reduced CA.

Nevertheless, the dispersed phases' insertion influenced the membrane's CA. The average values of CA in Figure 10c demonstrated that pure PBAT has CA = 55.9°. Botta and co-workers found a CA of 55° for PBAT membranes, like this work. However, the authors observed that the insertion of biochar into the PBAT matrix increased the CA, increasing hydrophobicity, ¹⁰³ unlike what occurred for the dispersed phases of this work, as seen in Figure 10c. The 0.5% C20A and/or NC loading was sufficient to decrease the CA. The EIPS/NIPS membranes with one dispersed phase had the lowest CA's, with PBAT/0.5 NC with CA = 46.5° and with prominence for PBAT/0.5 C20A with CA = 37.9°. Although both curcumin¹⁰⁴ and organomodified Cloisites¹⁰⁵ are hydrophobic, in the literature, it has already been discussed that both unmodified and modified MMTs, such as Cloisite, ¹⁰⁶⁻¹⁰⁷ and curcumin-based materials¹⁰⁸ can increase the hydrophilic character of polymeric films. Therefore, the insertion of nanoparticles may have increased

the surface roughness by increasing the surface capillary action and, consequently, the interaction with water. ¹⁰⁹ This behavior of composite membranes may favor the application of the membranes for adsorption and filtration. In the case of PBAT/0.5 C20A + 0.5 NC, the CA = 54.4° was similar to the pure PBAT membrane, showing that the interaction of this hybrid EIPS/NIPS membrane surface with the aqueous medium is negatively affected by the simultaneous action of the dispersed phases.

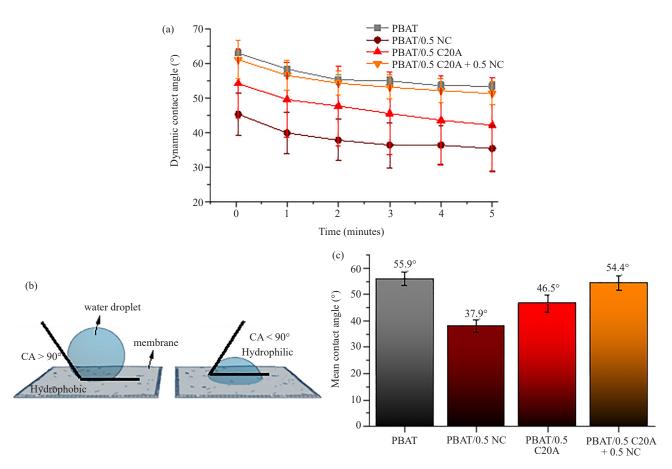


Figure 10. (a) Dynamic contact angle for the EIPS/NIPS membranes in the range of 0 to 5 min, (b) representation of a water droplet in a hydrophilic and a hydrophobic surface, and (c) mean values of the contact angle of the membranes

3.7 Copper adsorption

The adsorption test performed on the EIPS/NIPS membrane demonstrates the ability of the membrane to be an adsorbent material having a highly concentrated solution of contaminant, simulating the industrial effluent from a plating industry and the leak from the Brumadinho disaster (~ 1,000 mg/L). Figure 11 shows the sorption efficiency (%) aiming to decrease the initial concentration of 1,000 mg/L of Cu (II).

Figure 11 demonstrated that the EIPS/NIPS membrane could decrease the contaminated solution's initial concentration. The times studied showed variation in efficiency up to 24 h. The pure PBAT membrane, e.g., showed an efficiency of 20.5% in 10 min and 21.2% in 24 h, i.e., an increase of only 3.4%. The minor variation in sorption efficiency may be due to the saturation of the monolayer of the EIPS/NIPS membranes surface in contact with a solution of high Cu (II) concentration, reaching maximum adsorption.¹¹¹

The hybrid EIPS/NIPS membranes improved sorption efficiency, with membranes with only one dispersed phase being the most prominent in sorption efficiency at the end of the analysis (24 h), in the following order PBAT/0.5% NC (28.0%) > PBAT/0.5% C20A (27.0%) (Figure 11). Gaikwad and co-workers achieved a 27% removal efficiency of Cu

(II) from mining wastewater at pH 5.¹¹² In the case of the PBAT/0.5 C20A + 0.5 NC, no increase in adsorption efficiency was observed, having diminished their adsorption efficiency from T1 to T3 (24.4% to 16.7%, respectively). Thus, between 10 min and 24 h, there was a 31.5% reduction in efficiency. The literature addresses that the interaction of the material with the solution containing the PTE may initially saturate the adsorption sites and cause the sorption efficiency to be reduced by breaking bonds that interact with the contaminant. Yu and co-workers observed a 17% drop in Cu (II) sorption efficiency due to the breaking of active sites of chitosan in contact with contaminated industrial effluent. Karimi obtained a similar adsorption efficiency to this work using Ca(OH)₂, with a 20% removal efficiency of Pb (II) at pH 6 from industrial wastewater. Kumar and co-workers mentioned a 10% and 20% removal efficiency of Cu (II) and Ni (II) in contaminated water using Trien. The

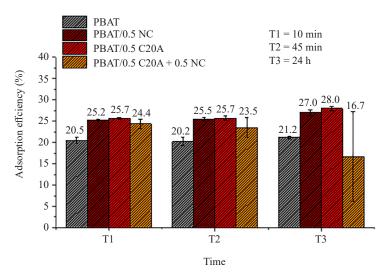


Figure 11. Sorption efficiency (%) of the adsorption tests of the membranes at times T1 = 10 min, T2 = 45 min, and T3 = 1,440 min

Therefore, the results are attractive since they illustrate the potential application of hybrid EIPS/NIPS membranes for removing PTEs, for which the inclusion of NC and C20A showed a potential to increase the membrane's potential. Additionally, using eco-friendly materials is appealing to reduce environmental impacts and promote sustainable treatment alternatives to increase water quality.

4. Conclusion

This work proposes the development of poly (butylene adipate-co-terephthalate) (PBAT) membranes with the inclusion of nanocurcumin (NC) and Cloisite® 20A (C20A) for the removal of toxic contaminants from water. Both dispersed phases presented nanometric size but with agglomeration effects observed due to their low ζ potential. C20A showed a crystallinity index (85.6%), while NC presented amorphization due to mechanical milling. The hybrid EIPS/NIPS membranes showed smaller pores, also impacting the membrane's mechanical performance, with the C20A promoting an increase in elastic modulus and tensile strength. In contrast, the PBAT/0.5 C20A + 0.5 NC showed reduced properties due to the amorphous nature of NC. EIPS/NIPS membranes present a hydrophilic nature, which is attractive for water contact applications. The sorption results showed that the membranes effectively reduced copper concentration even at low contact time, removing 25.7% for the PBAT/0.5 NC and 25.2 for PBAT/0.5 C20A. Thus, the PBAT/0.5 C20A + 0.5 NC membrane demonstrated a reduction in efficiency over time. The significant reduction in initial high concentration highlights the membrane's capability to eliminate toxic elements from water. Moreover, these findings emphasize using eco-friendly materials for water decontamination, providing a promising substitute for conventional treatment approaches. With the results obtained in this work, in future work in Cu (II) adsorption experiments using

C20A and NC in EIPS/NIPS membranes with PBAT matrix, other variables can be evaluated, such as adsorbent amount (NC and C20A alone or simultaneously) and Cu (II) solution concentration.

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Conflict of interest

The authors declare no competing financial interest.

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